

Rehabilitation of degraded areas in northeastern Patagonia, Argentina: Effects of environmental conditions and plant functional traits on performance of native woody species

Juan M ZEBERIO^{1,2*}, Carolina A PÉREZ³

¹ National University of Río Negro, Atlantic Headquarters, Center for Environmental Studies from Norpatagonia (CEANPa), Viedma 8500, Argentina;

² National Council of Scientific and Technical Research (CONICET), Viedma 8500, Argentina;

³ Ecological and Environmental Systems Research Laboratory (LISEA), National University of La Plata, La Plata 1900, Argentina

Abstract: Degradation processes affect a vast area of arid and semi-arid lands around the world and damage the environment and people's health. Degradation processes are driven by human productive activities that cause direct and indirect effects on natural resources, such as species extinction at regional scale, reduction and elimination of vegetation cover, soil erosion, etc. In this context, ecological rehabilitation is an important tool to recover key aspects of the degraded ecosystem. Rehabilitation trials rely on the use of native plant species with characteristics that allow them to obtain high survival and growth rates. The aim of this work was to assess the survival and growth of native woody species in degraded areas of northeastern Patagonia and relate them to plant functional traits and environmental variables. We observed high early and late survival rates, and growth rates in *Prosopis flexuosa* DC. var. *depressa* F.A. Roig and *Schinus johnstonii* F.A. Barkley, and low values in *Condalia microphylla* Cav. and *Geoffroea decorticans* (Gillies ex Hook. & Arn.) Burkart. Early survival rates were positively associated with specific leaf area (SLA) and precipitation, but negatively associated with wood density, the maximum mean temperature of the warmest month and the minimum mean temperature of the coldest month. Late survival rates were positively associated with SLA and soil organic matter, but negatively associated with plant height and precipitation. The temperature had a positive effect on late survival rates once the plants overcame the critical period of the first summer after they were transplanted to the field. *Prosopis flexuosa* and *S. johnstonii* were the most successful species in our study. This could be due to their functional traits that allow these species to acclimatize to the local environment. Further research should focus on *C. microphylla* and *G. decorticans* to determine how they relate to productive conditions, acclimation to environmental stress, auto-ecology and potential use in ecological rehabilitation trials.

Keywords: arid lands; *Condalia microphylla*; *Geoffroea decorticans*; *Prosopis flexuosa*; *Schinus johnstonii*; survival rates; height growth; basal diameter growth

*Corresponding author: Juan M ZEBERIO (E-mail: jmzeberio@unrn.edu.ar)

Received 2019-12-15; revised 2020-06-08; accepted 2020-06-23

© Xinjiang Institute of Ecology and Geography, Chinese Academy of Sciences, Science Press and Springer-Verlag GmbH Germany, part of Springer Nature 2020

1 Introduction

Degradation processes affect vast areas of arid and semi-arid lands around the world, which are inhabited by 2.5×10^9 people (UNCCD, 2011). Most of these lands are located in developing countries and display several degradation symptoms, affecting the health of people and ecosystems (James et al., 2013).

Ecosystem degradation is caused by human activities, evidenced by the loss of soil fertility and vegetation cover, soil erosion and the diminishing of land productivity (Abraham et al., 2009; Vallejo et al., 2009; Cortina et al., 2011). In this scenario, ecological rehabilitation is a key strategy to recover degraded ecosystems (Aronson et al., 1993; Maestre et al., 2001; Delgado-Baquerizo et al., 2013).

Ecological rehabilitation involves interventions on the natural system to return some of its lost structural and functional attributes. To achieve this objective, people should remove the disturbance factors and take actions to increase the vegetation cover (Aronson et al., 1993; Hobbs and Kramer, 2008). Some of the techniques used for improving vegetation cover involve sowing seeds of native or exotic species in bare soils, transplanting plants to the field, and/or applying soil amendments to improve the physical and chemical soil properties (Peláez et al., 1996; de Villalobos and Peláez, 2001; Maestre et al., 2001; Novak and Prach, 2003; de Villalobos et al., 2005; Milgrom, 2008; Dalmasso, 2010; Landi and Renison, 2010; Neri and Sánchez, 2010; Delgado-Baquerizo et al., 2013; Meli et al., 2013; Gasch et al., 2014; Soterias et al., 2014; Torres et al., 2015; Mcadoo et al., 2016; Plaza Behr et al., 2016; Pérez et al., 2019). Active plantation of species produced in plant nursery is one of the most frequently used techniques in rehabilitation trials (Pérez et al., 2019). In addition, it is desirable to initiate ecological rehabilitation processes using genetic material from the intervened area because native plant species are adapted to the dominant environmental conditions (Aronson et al., 1993; Cortina et al., 2006; Hobbs and Kramer, 2008).

Besides, the success of the rehabilitation project depends largely on the selected species (Cortina et al., 2006). The framework of functional traits has recently emerged as a useful tool to identify the most suitable species for rehabilitation programs (Martínez-Garza et al., 2013; Laughlin, 2014; Ostertag et al., 2015; Engst et al., 2016; Zirbel et al., 2017). This approach helps us to understand and predict the responses of species traits to certain environmental conditions and determine how these traits may affect ecosystem functions (Lavorel and Garnier, 2002; Zirbel et al., 2017). For example, wood density has shown a significant and positive relation with resistance to cavitation, and specific leaf area (SLA) has been related to competitive ability and stress tolerance. Species with a high SLA have faster growth rates and are less stress-tolerant (Westoby and Wright, 2006). Plant height is a predictor of depth root system and its ability to explore deep soil layers. Tall plants present deep roots, while dwarf shrubs present extended and shallow root systems (Weiher et al., 1999).

The study area is placed in northeastern Patagonia, Argentina. It is an ecotonal environment between the Monte and Espinal phytogeographic regions. From the structural point of view, it is a shrubby steppe with isolated trees surrounded by tall grasses and herbaceous species. The plant communities develop in a wide variety of soils and reliefs (Leon et al., 1998; Bran et al., 2000; Godagnone and Bran, 2009; Morello et al., 2012; Oyarzabal et al., 2018).

Northeastern Patagonian ecosystems present obvious signs of degradation and fragmentation (Zeberio, 2012, 2018; Abraham et al., 2016). The dominant soil use in this area is the production of extensive livestock and rainfed agriculture. Crops have been abandoned because of the loss of profitability caused by periodic drought (Zeberio et al., 2018). Ecological restoration studies in this area have different scopes and they are not always available (Bran et al., 2007; Gaitán et al., 2007; 2008; Kropfl et al., 2011; Funk et al., 2012; Peter et al., 2012; Zeberio, 2018). In addition, research on the performance of native woody species in rehabilitation projects in this area is particularly scarce (Torres Robles et al., 2015; Zeberio et al., 2018). To the best of our knowledge,

no previous studies have related the functional traits of these species to their performance in the harsh environmental conditions of degraded northeastern Patagonian.

We developed a structural equation model (SEM) based on hypothesized relationships among environmental conditions, plant functional traits and species responses. SEM is a useful tool for understanding the direct and indirect effects of predictors in complex multivariate systems (Grace et al., 2010). The aim of this work was, therefore, to assess the survival and growth of native woody species in degraded areas of northeastern Patagonia and to link these plant responses to plant functional traits and environmental variables.

2 Materials and methods

2.1 Site selection

We identified degraded areas in northeastern Patagonia based on geomorphological units, as defined by Godagnone and Bran (2009). According to Morello et al. (2012), the study area is located in the eco-region known as Monte de plains and plateaus.

We selected the sites to develop rehabilitation trials according to two criteria: (1) they have a common land-use history (Table 1), but (2) differential geomorphological characteristics representative of the high heterogeneity of the area (Table 2; Fig. 1). We selected three sites and named them after the landowners' last names. Erripa (Err) and Iturburu (Itu) are located in Adolfo Alsina district (Río Negro Province) and Bosero (Bos) in Patagones district (Buenos Aires Province). Degraded lands represent 40% of the total surface of these districts (Pezzola and Winschel, 2004; Zeberio, 2012).

Table 1 Location and land use in each rehabilitation trial site

Site	Latitude	Longitude	Land use	Age of clearing (a)
Err	41°09'11"S	63°25'19"W	Rainfed agriculture	25
Itu	41°01'04"S	63°01'40"W	Extensive livestock	23
Bos	40°39'30"S	62°52'00"W	Rainfed agriculture	20

Table 2 Characteristics of rehabilitation trial sites

Site	Geomorphological region	Soil texture	Depth (m)	pH	EC (dS/m)	BD (g/cm ³)	OM (mg/g)	TN (mg/g)	P (mg/g)
Err	Loess plateau	Sandy	>1.00	8.2	0.58	1.27	0.53	0.06	0.0089
Itu	Loess plateau	Sandy-loamy	0.45	8.2	0.54	1.41	1.06	0.09	0.0087
Bos	Interfluvial plateau	Loamy	0.12	7.8	1.78	0.43	0.30	0.03	0.0023

Note: EC, electrical conductivity; BD, bulk density; OM, organic matter; TN, total nitrogen; P, phosphorus.

Three 100 m²-plots were established in each site. The plots were at least 75 m apart from each other and fenced with a 60-cm-high plastic mesh to actively protect planted species against potential damage caused by livestock and small mammal herbivores.

The climate data were obtained from WorldClim database (Fick and Hijmans, 2017). The mean annual precipitation during 2013–2016 presented a bimodal behavior, with two maxima. Rainfall reached a peak in the three sites in April, a second peak in September at Bos and in October in the other two sites. Accumulated annual precipitations at Bos, Itu, and Err sites were 443.0, 460.3, and 455.3 mm, respectively. In autumn and winter, the highest precipitation was recorded at Bos and Err sites. In the spring and summer, the highest precipitation was observed at Err, the site closest to the sea (Fig. 2). The maximum mean temperature for January (the warmest month) ranged from 27.2°C to 25.8°C, and the minimum mean temperature for July (the coldest month) ranged from 0.6°C to –0.2°C, depending on the site.

The soil chemical properties were estimated from a composite soil sample (30 cm in depth) and collected from the center of each plot. We estimated soil organic matter (OM) by the Walkley and Black method, total nitrogen (TN) by the micro-Kjeldahl method, and total phosphorus (P) by the

Olsen method. Further, we registered soil depth in each site until lithic contact and extracted a core to determine soil density (de Inalbon, 2005).

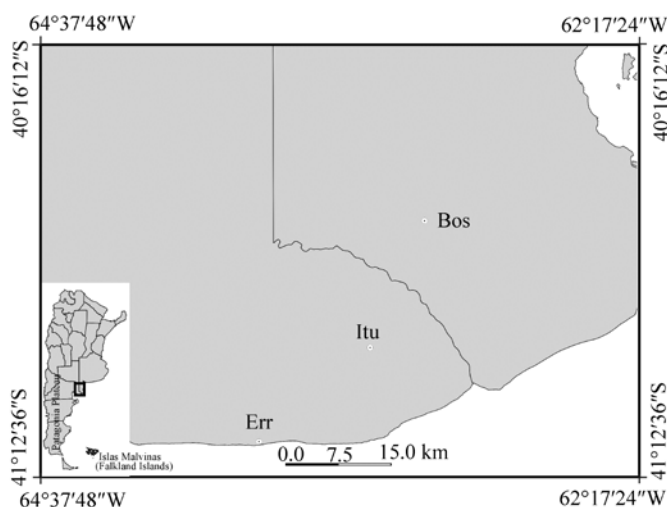


Fig. 1 Location of rehabilitation trial sites (Bos, Itu and Err) in northeastern Patagonia

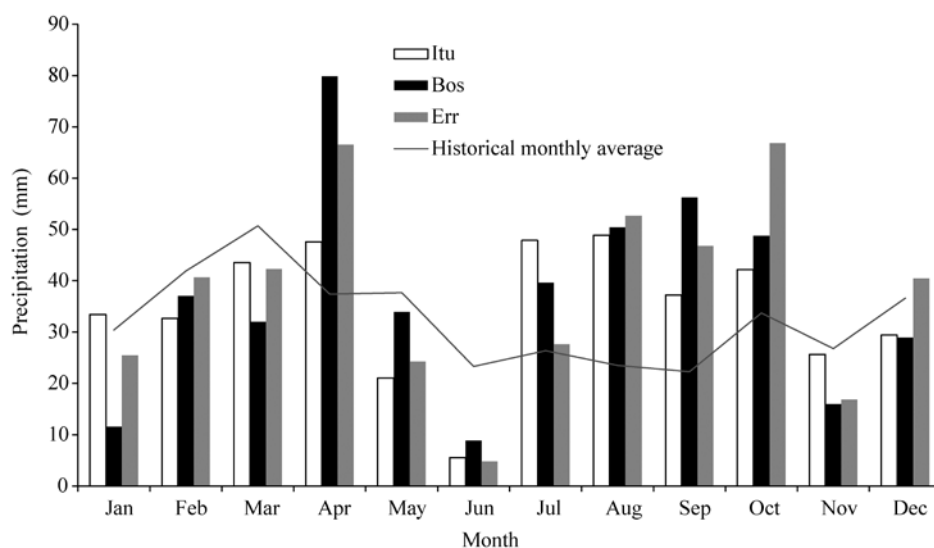


Fig. 2 Precipitation in rehabilitation trial sites (Bos, Itu and Err) and historical monthly average precipitation during 2013–2016. Data were cited from de Berasategui (2004).

2.2 Species selection

We selected the most frequent woody species in non-degraded areas (Zeberio et al., 2018), namely the shrubs *Prosopis flexuosa* DC. var. *depressa* F.A. Roig, *Condalia microphylla* Cav. and *Schinus johnstonii* F. A. Barkley, as well as the tree *Geoffroea decorticans* (Gillies ex Hook. & Arn.) Burkart (Table 3).

Seeds of the selected species were collected from the study area in order to produce individuals in a greenhouse. The seeds were collected from plants that met the selection criteria by Ffolliott and Thames (1983) and stored in paper bags until sowing. The seeds were sown between May and July 2012 in the greenhouse of experimental station of the National Institute of Agricultural Technology, Argentina.

The most adequate pre-germination treatments were conducted for each species. *Condalia microphylla*, *G. decorticans* and *P. flexuosa* seeds were scarified with sulphuric acid (98% P/V) for 32 min. In the case of *S. johnstonii*, we removed the exo- and mesocarp of the seeds (Zeberio

and Calabrese, 2013). The seeds were sown in plastic trays of 60 cm×48 cm×15 cm containing a mixed substrate with equal parts of sand and peat to facilitate seedling removal. A month later, the seedlings were placed in individual pots.

Table 3 Summary of plant characteristics used in rehabilitation trial sites

Woody species	SLA (cm ² /g)	WD (g/cm ³)	Foliage	Height (m)	Life form
<i>C. microphylla</i>	26.5	1270	Evergreen	0.5–2.0	Shrub
<i>G. decorticans</i>	103.2	970	Deciduous	1.4–4.0	Tree
<i>S. johnstonii</i>	28.5	995	Evergreen	1.5	Shrub
<i>P. flexuosa</i>	141.5	980	Deciduous	2.5	Shrub

Note: SLA, specific leaf area; WD, wood density.

Each individual pot contained a substrate of soil extracted from the experimental sites (50%), commercial earthworm (25%), and peat (25%). The plants were grown in the greenhouse for about seven months with periodic irrigation (twice a week). After that, an acclimation process was developed (Cortina et al., 2006). A month before the plants were transplanted to the field, they were gradually exposed to natural environmental conditions, including direct exposure to sunlight and wind, as well as irrigation decrease (once a week).

2.3 Experimental design

The plants were transplanted to the experimental sites between April and May 2013. For each species, nine individuals were planted in two rows, except for *C. microphylla* (five plants in one row). This species presented high mortality in the acclimation phase and we could not obtain enough individuals. The rows were arranged 2 m apart from each other. The plant holes were 15 cm in diameter and 20 cm in depth, and 1200 cm³ of commercial earthworm was added at the bottom of the holes to prevent compaction. The remaining spaces in the holes were filled with the soil extracted when digging the plant hole and manually compacted. To improve moisture retention, we dug a basin around each plant to capture and retain natural moisture.

In December 2013, we recorded the survival of individuals of each species (early survival rate), plant height and basal diameter. Three years later, in February 2016, we recorded the late survival rate and measured plant height and basal diameter of each alive plant. The growth of each plant was estimated as the difference between the initial and final measurements of height and basal diameter. Plant functional traits (SLA, wood density and plant height) were obtained from TRY (not an acronym, rather an expression of sentiment: <http://www.try-db.org>) plant trait database (Kattge et al., 2011) (Table 3).

2.4 Data analyses

Generalized linear models were performed to evaluate plant survival and growth. Binomial distribution of survival data and normal distribution of growth data were assumed. Site and species were considered predictor variables. Site-species interaction was estimated to evaluate the different performance of the species in each site. We compared the means using the LSD (least significant difference) test. Data were analyzed by Infostat software (di Rienzo et al., 2008).

The causal relationships between environmental variables and plant functional traits were analyzed to determine their effects on early survival rates, late survival rates, plant height and basal diameter. Environmental variables and plant functional traits were selected by linear correlation with plant survival and growth rates. Structural equation models (SEM) are a synthesis for path analysis, factorial analysis and the maximum likelihood techniques. They are used in ecological science as causal inference. This analysis can prove the plausibility of causal models based on a priori information on the relationship between the variables of interest. We used SEM to evaluate the relative importance and direct/indirect effects of climate, soil variables and plant functional traits on plant survival and growth. We used the normed fit index and the root mean square error of approximation index as measures of model fit to our data (Grace, 2006). Path

coefficient estimates were obtained using the maximum-likelihood estimation technique, which are the equivalent of standardized partial regression coefficients and are interpreted as relative effects of one variable upon another (Grace et al., 2010).

3 Results

3.1 Early survival rate

Early survival rate was estimated seven months after the plants were transplanted to the restoration trial sites. *Prosopis flexuosa*, *G. decorticans* and *S. johnstonii* presented high survival rates (near 100%) ($n=288$; $df=6$; $F=124.4$; $P<0.001$). *Condalia microphylla* presented lower survival rate (50%) than the other species ($P<0.001$) (Fig. 3). Bos and Err sites showed higher survival rates (85% and 95%, respectively) than Itu (75%) ($P=0.046$). Species \times sites interaction of the early survival rate was significant only for *C. microphylla* ($P=0.09$). This species presented higher survival rate at Err (70%), but lower at Bos (50%) and Itu sites (20%).

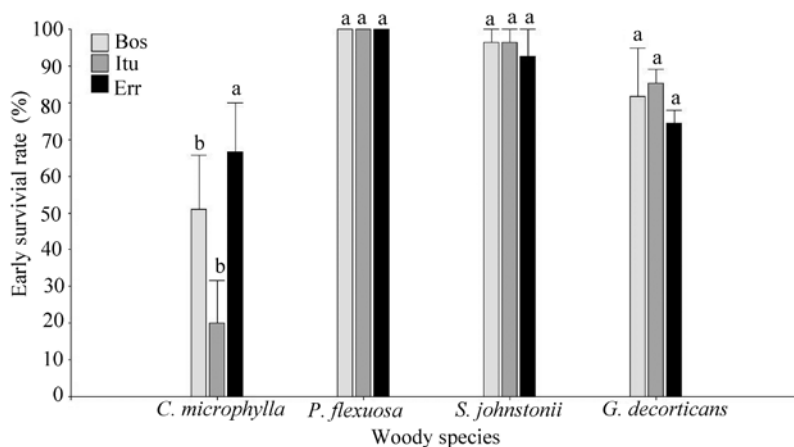


Fig. 3 Woody species early survival rate recorded in December 2013 (seven months after field planting). Different lowercase letters indicate significant differences among different sites and woody species at $P<0.05$ level.

3.2 Late survival rate

Late survival rate was recorded three years after the transplantation date, in the summer of 2016. Survival rates decreased with time, as late survival rate was smaller than early survival rate ($n=224$; $df=6$; $F=296.6$; $P<0.001$). *Prosopis flexuosa* and *S. johnstonii* presented higher survival rates (70% and 75%, respectively) than *C. microphylla* and *G. decorticans* (15% and 35%, respectively) ($P=0.0003$) (Fig. 4). Late survival rate at Itu (60%) was higher than those at Bos and Err sites (42% and 38%, respectively) ($P=0.0074$). Species \times site interaction was not significant ($P=0.186$). The transplanted specimens of *G. decorticans* were severely browsed by the Patagonian hare (*Dolichotis patagonum*), and their stems were felled at ground level.

3.3 Species growth

Three years after transplantation, plant height ranged from 25.50 cm (*C. microphylla*) to 38.54 cm (*S. johnstonii*) ($n=224$; $df=6$; $F=113.0$; $P<0.001$). The height increase of *P. flexuosa* (9.76 cm) and *S. johnstonii* (9.58 cm) was higher than that of *G. decorticans* (5.99 cm). *Condalia microphylla* was the species with the smallest height increase (1.04 cm; $P=0.0002$) (Fig. 5). The increase in plant height was higher at Itu than at Bos and Err sites ($P<0.0001$). Species \times site interaction was significant ($P=0.047$) because *G. decorticans* and *S. johnstonii* presented a higher height increase at Itu than at the other sites.

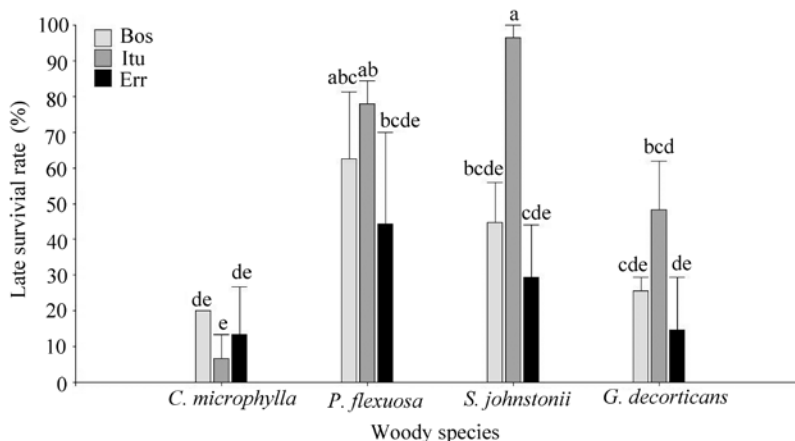


Fig. 4 Woody species late survival rate recorded in February 2016 (thirty one months after field planting). Different lowercase letters indicate significant differences among different sites and woody species at $P < 0.05$ level.

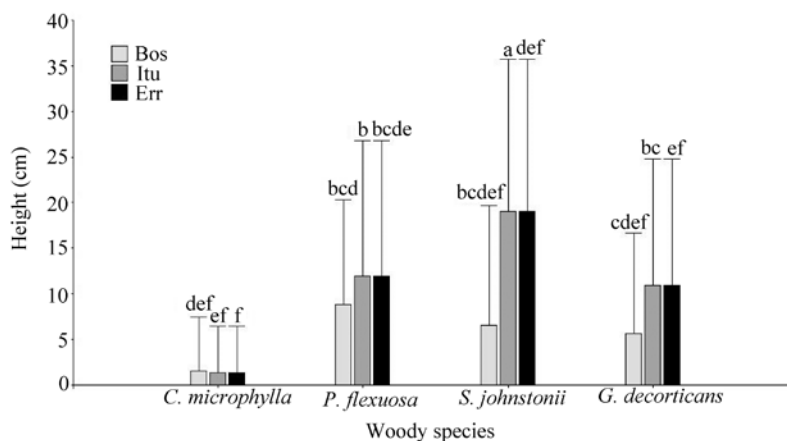


Fig. 5 Woody species height during 2013–2016. Different lowercase letters indicate significant differences among different sites and woody species at $P < 0.05$ level.

The basal diameter of stem varied from 0.47 (*G. decorticans*) to 0.96 cm (*S. johnstonii*) ($n=288$; $df=6$; $F=125.6$; $P < 0.001$). The highest increase of plant basal diameter was found for *S. johnstonii* (0.29 cm) and *P. flexuosa* (0.21 cm) ($P < 0.001$) (Fig. 6). Itu presented a higher basal diameter increase than the other two sites ($P < 0.001$). Species \times site interaction showed significant differences ($P < 0.001$). *Schinus johnstonii* showed a higher increase in basal diameter at Itu site, but *C. microphylla*, *G. decorticans* and *P. flexuosa* presented no differences among sites.

3.4 Effects of environmental variables and plant functional traits on survival and growth

A structural equation model was built to evaluate the relations between environmental variables and functional plant traits over earlier survival rate, late survival rate and growth (chi-square=43.5; $P=0.18$; $df=28$; normed fit index=0.87). This model explained 91% of the variation in early survival rate, height growth and diameter growth, and 95% of late survival rate (Fig. 7). The graphs show only the positive and negative statistically significant effects ($P < 0.05$). Wood density was negatively associated with late survival rate. SLA was positively associated with both early and late survival rates. The common height of this species had a negative effect on late survival rate (Fig. 7).

The minimum and maximum temperatures had a negative effect on early survival rate but a positive effect on late survival rate. The accumulated seasonal precipitation in the study period

positively affected early survival rate but negatively affected late survival rate and height growth (Fig. 7).

Soil OM and soil depth to lithic contact were the edaphic variables that showed statistical significance in our model. Soil OM had a positive effect on late survival rate, while soil depth had a negative effect on early survival rate and height growth (Fig. 7).

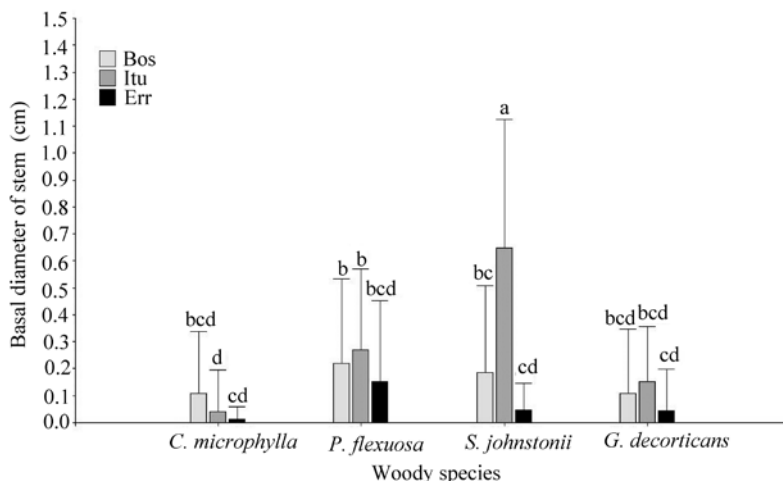


Fig. 6 Woody species basal diameter of stem during 2013–2016. Different lowercase letters indicate significant differences among different sites and woody species at $P < 0.05$ level.

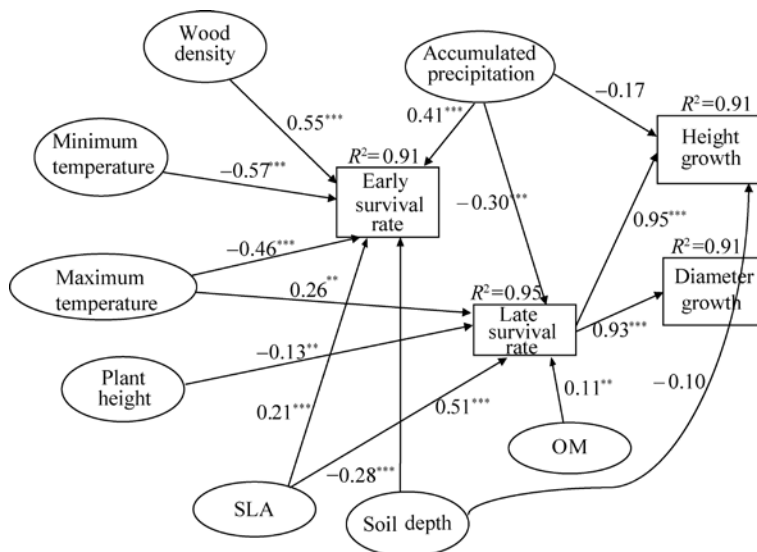


Fig. 7 Structural equation model (SEM) for the environmental variables (minimum temperature in July, maximum temperature in January, soil depth and soil organic matter (OM), specific leaf area (SLA), wood density and plant height) and estimated variables (early survival rate, late survival rate, diameter growth and height growth). The numbers next to arrows are path coefficients and they show the strengths of the effect. Non-significant ($P > 0.05$) paths were eliminated. ** and *** indicate significances among variables at $P < 0.01$ and $P < 0.001$ levels, respectively.

4 Discussion

In this study, the selected species showed high early and late survival rates related to the prevailing strategy in arid and semi-arid environments. Two contrasting plant strategies have been identified. One of them is the acquisitive strategy, involving a set of functional attributes that facilitates the rapid acquisition of resources and high growth rates. The other one is the

conservative strategy, involving another set of attributes that allows the plant to keep the acquired resources, but at lower growth rates (Poorter, 1990; Lambers and Poorter, 1992; Reich et al., 1997; Aerts and Chapin, 2000; Wright and Westoby, 2002; Díaz et al., 2004). SLA values of the woody species used in this study had a positive impact on their early and late survival rates. Plant species with low SLA values tend to have sclerophyllous leaves. Their thick epidermal walls keep up moisture more efficiently and have lower canopy height, and small xylem vessels that prevent cavitation and high specific wood density (Wright and Westoby, 2002; Díaz et al., 2004; Westoby and Wright, 2006). The prevailing plant strategies are associated with the environmental characteristics of the region in which they are established (Díaz et al., 2004). In the arid and semi-arid regions of Argentina, the species have relatively low SLA values (Díaz et al., 1999). This is the case for the species in our study because their leaves have anatomical structures that protect the photosynthetic tissues from desiccation and herbivory, such as thick cuticles, presence of spines and anti-herbivore compounds (Peláez et al., 1994; Kraus et al., 2003; Bucci et al., 2004; Villagra et al., 2011). Therefore, these species may be considered conservative according to their low SLA values. These conservative strategies could have helped the selected species to achieve high or moderate survival rates in the harsh environmental conditions in the study sites.

SLA negatively correlated with wood density. Wood density is considered a proxy of vulnerability to embolism, which can endanger the growth and survival of plants in water stress conditions. Higher wood density is related to greater resistance to cavitation (Westoby and Wright, 2006; Chave et al., 2009). Thus, in our study, we expected a positive relationship between wood density and survival rate. However, wood density was negatively associated with early survival rate. We believed that this result was conditioned by the high mortality rate presented by *C. microphylla*, which had the highest wood density among the species used in this study. The high mortality of *C. microphylla* would be related to their ecophysiological characteristics. Peláez et al. (1996) reported low *C. microphylla* survival rate (25%) in the semi-arid region of Argentina, where the rainfall shortage impaired the establishment of transplanted individuals. In our work, it is likely that in addition to environmental constraints, the older age of the transplanted individuals also harmed their survival rates. The specimens of *C. microphylla* used in this rehabilitation trial presented a high mortality rate in the acclimation phase (unpublished data), and the older specimens had to be used to complete the required number of plants. Cortina et al. (2006) indicated that in order to obtain the high survival rates, the specimens of woody species destined for ecological restoration should be transplanted in their first vegetative stage.

The maximum and minimum temperatures may be indicators of the effects of extreme temperature on plants. These temperatures presented a negative effect on early survival rate of the plant species in this study. Instead, temperature showed a positive effect on late survival rate. These results would be associated with the critical threshold that plants must overcome during the first year after the transplantation to the field and the acclimation process (Cortina et al., 2006; Maestre et al., 2006).

Recorded rainfall had a positive effect on early survival rate but a negative effect on late survival rate and plant height in all transplanted individuals in our study. Godagnone and Brand (2009) indicated tendencies about rainfall were coincident with that of our precipitation record. Northeastern Patagonia is characterized by inter-annual and intra-annual precipitation variability (Godagnone and Bran, 2009). For the period measured in this study, the rainiest seasons were autumn (80 mm) and spring (71 mm), and the lowest amount of rainfall was recorded in the summer (25 mm). Early survival of plants was estimated in spring, one of the rainiest periods, but three years later, plants were exposed to summer rainfall shortages. Plant transpiration demands are highest during the summer, and the drought could have negatively affected the late survival rate and limited plant growth.

Soil OM values obtained of each study site (0.46 to 1.63 mg/g) was coincident with that reported by Godagnone and Bran (2009) for northeastern Patagonia. However, low soil OM values had a positive effect on late survival rate of plant species in our study. This might be related to the adaptation of native species, and their ability to obtain scarce soil resources and use them efficiently. Besides, the sites in this study presented a common land-use history, which was rainfed agriculture. Grman et al. (2013) found that residual nutrients from previous cropland use may have a positive effect on the survival and establishment of plants in restoration areas.

The common height of the plant species used in this study negatively affected late survival rate. This is linked to the lowest survival rate of *G. decorticans*, the only tree in this study. The plant height growth was related to soil depth. Schenk and Jackson (2005) indicate that woody plant height is a good predictor for the root system depth. When root systems are limited by some physical soil impediment, plants have a lower development of above-ground organs too (Peláez et al., 1994; Bucci et al., 2004). Bos site had the lowest soil depth due to a sub-superficial calcium carbonate layer, which acts as a barrier to root system development. The differences in plant height growth registered in the study sites indicate that soil depth had a sharp effect on the transplanted plants, regardless of their species identity. The height growth was higher in plants grown in deeper soil sites.

In our study, *P. flexuosa* and *S. johnstonii* were the most successful species, because they exhibited the highest survival and growth rates. Both species have a dimorphic root system. They not only have deep roots that allow them to uptake water from deep soil layers, but also have shallow roots to take the soil surface water supplied by occasional rainfall (Villagra, 2000; Bucci et al., 2009; Jobbágy et al., 2011; Villagra et al., 2011). The high early survival rate of *P. flexuosa* obtained in this study was similar to that obtained by López Launstein et al. (2012) in northwestern Argentina. In that study, *P. flexuosa* showed 100% of early survival rate under moderate water stress and 70% of survival rate under severe water stress in the field. In northwestern Patagonia, Pérez (2013) found a lower survival rate in *S. johnstonii* (75%) than that obtained in our study. Both species, *P. flexuosa* and *S. johnstonii* had different SLA values, but similar wood density and root system morphology. Dimorphic root systems are absent in the other species of this rehabilitation trial. *Condalia microphylla* presents a shallow radical system, extended in the first 60 cm of soil depth, which makes it compete with grasses to obtain rainfall water (Peláez et al., 1994). Instead, *G. decorticans* presents a pivoting and deep root system (Kraus et al., 2003), which makes it less able to take advantage of superficial soil water. The dimorphic root system of *P. flexuosa* and *S. johnstonii* allows them to obtain water from superficial and deep soil layers. This may help to understand the persistence of these species under water stress conditions (Díaz and Cabido, 1997; Whitford, 2002). In addition, *P. flexuosa* and *S. johnstonii* are shrubs, while *G. decorticans* is a tree. As *G. decorticans* was severely damaged by Patagonian hare in the field, we could not obtain the conclusive data.

5 Conclusions

The structural equation models used in this study related the performance of native woody species to functional traits and environmental variables in an ecological restoration trial in deforested semi-arid lands. *Prosopis flexuosa* and *S. johnstonii* were the most successful species in our study. These species overcome rainfall shortages and high-temperature stress. The presence of a dimorphic root system may allow them to have competitive advantages regardless of other functional traits. These advantages could be related to the use of both deep soil water and water of shallower soil layers.

Due to the damage caused by Patagonian hare, future field studies should focus on the survival and growth of *G. decorticans* to evaluate their inclusion in ecological rehabilitation projects in deforested areas of northeastern Patagonia. Similarly, as *C. microphylla* is a frequent species in northern Patagonia, auto-ecology studies should be conducted to determine whether this species may be included in rehabilitation projects.

Acknowledgements

This work was funded by the National University of Río Negro, Argentina (PI40c658, PI40c654). We thank the staff and owners of the study sites Bosero, Iturburu and Erripa. Two anonymous reviewers greatly improved the quality of this work.

References

- Abraham M E, Corso M L, Maccagno P. 2011. Drylands and desertification in Argentina. In: Desertification Assessment in Argentina. LADA/FAO Project Results, 11–64. (in Spanish)

- Abraham M E, Guevara J C, Candia R, et al. 2016. Dust storms, drought and desertification in the southwest of Buenos Aires Province, Argentina. *Faculty Journal of Agricultural Sciences of Cuyo*, 48 (2): 221–241.
- Aerts R, Chapin F S. 1999. The mineral nutrition of wild plants revisited: a re-evaluation of processes and patterns. *Advances in Ecological Research*, 30: 1–67.
- Aranson J, Floret C, Floc'h E, et al. 1993. Restoration and rehabilitation of degraded ecosystems in arid and semiarid lands. A view from the south. *Restoration Ecology*, 1(1): 8–17.
- Bran D, Lopez C, Auesa, J. 2000. Ecological Regions of Río Negro. In: National Institute of Agricultural Technology. Editorial INTA, Bariloche, Argentina. (in Spanish)
- Bran D E, Cecchi G, Gaitán J J. 2007. Effect of burn severity on vegetation recovery in the Monte Austral. *Ecología Austral*, 17(1): 123–131. (in Spanish)
- Bucci S, Goldstein G, Meinzer F, et al. 2004. Functional convergence in hydraulic architecture and water relations of tropical savanna trees: from leaf to whole plant. *Tree Physiology*, 24(8): 891–899.
- Bucci S, Scholz F, Goldstein G, et al. 2009. Soil water availability and rooting depth as determinants of hydraulic architecture of Patagonian woody species. *Oecologia*, 160(4): 631–641.
- Chave J, Coomes D, Jansen S, et al. 2009. Towards a worldwide wood economics spectrum. *Ecology Letters*, 12(4): 351–366.
- Cortina J, Maestre F, Vallejo R, et al. 2006. Ecosystem structure, function, and restoration success: Are they related? *Journal for Nature Conservation*, 14(3): 152–160.
- Cortina J, Amat B, Castillo V, et al. 2011. The restoration of vegetation cover in the semi-arid Iberian southeast. *Journal of Arid Environments*, 75(12): 1377–1384.
- Dalmaso D A. 2010. Revegetation of degraded areas with native species. *Boletín de la Sociedad Argentina de Botánica*, 45(1–2): 149–171. (in Spanish)
- de Berasategui L. 2004. Climatic Statistics of the Viedma Valley–30 Years. Sáenz Peña: National Institute of Agricultural Technology, 7–62. (in Spanish)
- de Inalbon M R, Valenzuela A. 2005. Analytical Procedures for Normal and Saline Soils. Sáenz Peña: National Institute of Agricultural Technology, 4–28. (in Spanish)
- de Villalobos A E, Peláez D V. 2001. Influences of temperature and water stress on germination and establishment of *Prosopis caldenia* Burk. *Journal of Arid Environments*, 49(2): 321–328.
- de Villalobos A E, Peláez D V, Elía O. 2005. Growth of *Prosopis caldenia* Burk: seedlings in central semi-arid rangelands of Argentina. *Journal of Arid Environments*, 61(3): 345–356.
- Delgado-Baquerizo M, Maestre F, Gallardo A, et al. 2013. Decoupling of soil nutrient cycles as a function of aridity in global drylands. *Nature*, 502(73–74): 670–672.
- di Rienzo J, Casanoves F, Balzarini M, et al. 2008. InfoStat. Version 2008. Córdoba: National University of Córdoba. Argentina. (in Spanish)
- Díaz S, Cabido M. 1997. Plant functional types and ecosystem function in relation to global change. *Journal of Vegetation Science*, 8(4): 463–474.
- Díaz S, Cabido M, Zak M, et al. 1999. Plant functional traits, ecosystem structure and land-use history along a climatic gradient in central- western Argentina. *Journal of Vegetation Science*, 10(5): 651–660.
- Díaz S, Hodgson J G, Thompson K, et al. 2004. The plant traits that drive ecosystems: Evidence from three continents. *Journal of Vegetation Science*, 15(3): 295–304.
- Engst K, Baassch A, Erfmeier A, et al. 2016. Functional community ecology meets restoration ecology: Assessing the restoration success of alluvial floodplain meadows with functional traits. *Journal of Applied Ecology*, 53(3): 751–764.
- Ffolliott, P F, Thames J L. 1983. Collection, Handling, Storage and Pre-treatment of *Prosopis* Seeds in Latin America. Rome, Italy: Food and Agriculture Organization of the United Nations, 7–45.
- Fick S E, Hijmans R J. 2017. WorldClim 2: New 1-km spatial resolution climate surfaces for global land areas. *International Journal of Climatology*, 37(12): 4302–4315.
- Funk F, Peter G, Loydi A, et al. 2012. Structural and functional recovery of intershrub spaces after 10-years of grazing exclusion in a semiarid steppe of northeastern Patagonia. *Ecología Austral*, 22(3): 195–202. (in Spanish)
- Gaitán J J, Bran D E, Murray F. 2007. Effect of burn severity on the soil organic carbon concentration mounds and intermounds in the Monte Austral. *Soil Science*, 25(2): 191–199. (in Spanish)
- Gaitán J J, López C R, Bran D E. 2008. Grazing effects on soil and vegetation in the Patagonian steppe. *Ecología Austral*, 27(2): 1–10.
- Gasch C, Huzurbazar S, Stahl P. 2014. Measuring soil disturbance effects and assessing soil restoration success by examining distributions of soil properties. *Applied Soil Ecology*, 76: 102–111.

- Godagnone R E, Bran D E. 2009. Integral Inventory of Natural Resources of the Province of Rio Negro. Buenos Aires: INTA, 319–363. (in Spanish)
- Grace J. 2006. Structural Equation Modeling and Natural Systems. Cambridge: Cambridge University Press, 207–275.
- Grace J, Anderson M, Olff H, et al. 2010. On the specification of structural equation models for ecological systems. *Concepts and Synthesis*, 80(1): 67–87.
- Grman E, Bassett, T, Brudvig L A. 2013. Confronting contingency in restoration: management and site history determine outcomes of assembling prairies, but site characteristics and landscape context have little effect. *Journal of Applied Ecology*, 50: 1234–1243.
- Hobbs R, Cramer V. 2008. Restoration ecology: Interventionist approaches for restoring and maintaining ecosystems function in the face of rapid environmental change. *Annual Review of Environment and Resources*, 33(1): 39–61.
- James J J, Svejcar T J, Rinella M J. 2011. Demographic processes limiting seedling recruitment in arid grassland restoration. *Journal of Applied Ecology*, 48(4): 961–969.
- Jobbágy E, Nosoetto M, Villagra P, et al. 2011. Water subsidies from mountains to deserts: Their role in sustaining groundwater-fed oases in a sandy landscape. *Ecological Applications*, 21(3): 678–694.
- Kattge J, Díaz S, Lavorel S, et al. 2011. TRY—a global database of plant traits. *Global Change Biology*, 17(9): 2905–2935.
- Kraus T, Bianco C, Weberling F. 2003. Root system morphology of Fabaceae species from central Argentina. *Wulfenia*, 10(2003): 61–72.
- Kropfl A I, Cecchi G, Villasuso N, et al. 2011. Degradation and recovery processes in semi-arid patchy rangelands of northern Patagonia, Argentina. *Land Degradation and Development*, 24(4): 393–399.
- Lambers H, Poorter H. 1992. Inherent variation in growth rate between higher plants: A search for physiological causes and ecological consequences. *Advances in Ecological Research*, 23: 187–261.
- Landi M, Renison D. 2010. Forestation with *Polylepis australis* BITT. in eroded soils of the Sierras Grandes of Córdoba. *Ecología Austral*, 20(1): 47–55. (in Spanish)
- Laughlin D. 2014. The intrinsic dimensionality of plant traits and its relevance to community assembly. *Journal of Ecology*, 102(1): 186–193.
- Lavorel S, Garnier E. 2002. Predicting changes in community composition and ecosystem functioning from plants traits: Revisiting the holy grail. *Functional Ecology*, 16(5): 545–556.
- Leon R, Bran D E, Collantes M, et al. 1998. Main vegetation units of the extra Andean Patagonia. *Ecología Austral*, 8: 125–144. (in Spanish)
- López Launstein D, Fernández M E, Verga A. 2012. Differences in drought responses of seedlings of *Prosopis chilensis*, *P. flexuosa* and interspecific hybrids: implications for reforestation in arid zones. *Ecología Austral*, 22(1): 43–52. (in Spanish)
- Maestre F, Bautista S, Cortina J, et al. 2001. Potential for using facilitation by grasses to establish shrubs on a semiarid degraded steppe. *Ecological Application*, 11(6): 1641–1655.
- Maestre F T, Martín N, Diez B, et al. 2006. Watering, fertilization, and slurry inoculation promote recovery of biological crust function in degraded soils. *Microbial Ecology*, 52(3): 365–377.
- Martínez-Garza C, Bongers F, Poorter L. 2013. Are functional traits good predictors of species performance in restoration plantings in tropical abandoned pastures? *Forest Ecology and Management*, 303: 35–45.
- Mcadoo J, Swanson J, Murphy P, et al. 2016. Evaluating strategies for facilitating native plant establishment in northern Nevada crested wheatgrass seedlings. *Restoration Ecology*, 25(1): 53–62.
- Meli P, Martínez-Ramos M, Rey-Benayas J, et al. 2013. Selecting species for passive and active riparian restoration in southern Mexico. *Restoration Ecology*, 21(2): 163–165.
- Milgrom T. 2008. Environmental aspects of rehabilitating abandoned quarries: Israel as a case study. *Landscape and Urban Planning*, 87: 172–179.
- Morello J, Mateucci S, Rodríguez A F, et al. 2012. Ecoregions and Argentine Ecosystem Complexes. Buenos Aires: Graphic Orientation, 309–347. (in Spanish)
- Neri A, Sánchez L. 2010. A procedure to evaluate environmental rehabilitation in limestone quarries. *Journal of Environmental Management*, 91(11): 2225–2237.
- Novak J, Prach K. 2003. Vegetation succession in basalt quarries: Pattern on a landscape scale. *Applied Vegetation Science*, 6(2): 111–116.
- Ostertag R, Warman L, Cordell S, et al. 2015. Using plant functional traits to restore Hawaiian rainforest. *Journal of Applied Ecology*, 52(4): 805–809.
- Oyarzabal M, Clavijo J, Oakley L, et al. 2018. Vegetation units of Argentina. *Ecología Austral*, 28(1): 40–63. (in Spanish)
- Peláez D V, Distel R, Bóo R, et al. 1994. Water relations between shrubs and grasses in semi-arid Argentina. *Journal of Arid Environments*, 27(1): 71–78.

- Peláez D V, Bóo R, Elfa O. 1996. The germination and seedling survival of *Condalia microphylla* Cav. in Argentina. *Journal of Arid Environments*, 32(2): 173–179.
- Pérez D. 2013. Restoration of arid and semi-arid Patagonian ecosystems. Implementation from an ecological and social perspective. In: Pérez D, Rovere A, Rodríguez Araujo M E. *Ecological Restoration in the Arid Diagonal of Argentina*. Buenos Aires: Vazquez Mazzini, 49–60. (in Spanish)
- Pérez D, Gonzáles F, Ceballos C, et al. 2019. Direct seeding and out plantings in drylands of Argentinean Patagonia: estimated costs, and prospects for large-scale restoration and rehabilitation. *Restoration Ecology*, 27(5): 1–12.
- Peter G, Funk F, Loydi A, et al. 2012. Variation in specific composition and cover in grassland exposed to various grazing pressures in the Monte Rionegrino. *Phyton*, 81: 233–237. (in Spanish)
- Pezzola A, Winschel C. 2004. Multitemporal study of the degradation of native forests in the Patagones district, Buenos Aires. In: Technical Report INTA EEA Ascasubi, Buenos Aires, Argentina. (in Spanish)
- Plaza Behr M, Pérez C, Goya J, et al. 2016. *Celtis ehrenbergiana* planting as a technique for the recovery of native forests invaded by *Ligustrum lucidum* in NE Buenos Aires. *Ecología Austral*, 26(2): 171–177. (in Spanish)
- Poorter H. 1990. Interspecific variation in relative growth rate: on ecological causes and physiological consequences. In: Lambers H, Cambridge M L, Konings H, et al. *Causes and Consequences of Variation in Growth Rate and Productivity of Higher Plants*. The Hague: SPB Academic Publishing, 45–68.
- Reich P B, Walters M B, Ellsworth D S. 1997. From tropics to tundra: Global convergence in plant functioning. *Proceedings of the National Academy of Sciences of the United States of America*, 94(25): 13730–13734.
- Schenk H J, Jackson R B. 2005. Mapping the global distribution of deep roots in relation to climate and soil characteristics. *Geoderma*, 126: 129–140.
- Soteras F, Renison D, Becerra A. 2014. Restoration of high altitude forests in an area affected by a wildfire: *Polylepis australis* Bitt. Seedlings performance after soil inoculation. *Trees Structure and Function*, 28(1): 173–182.
- Torres R, Giorgis M, Trillo C. 2015. Outplanting survival and growth of species with different life histories in burned and unburned sites: a case study of two woody species in the Chaco Serrano, Argentina. *Ecología Austral*, 25(2): 135–143. (in Spanish)
- Torres Robles S S, Arturi M F, Contreras C, et al. 2015. Geographical variations in structure and composition of woody vegetation in the border Espinal and Monte in Northeastern Patagonia. *Boletín de la Sociedad Argentina de Botánica*, 50(2): 209–215. (in Spanish)
- United Nations Convention to Combat Desertification (UNCCD), 2011. Desertification. A Visual Synthesis. [2019-07-07]. <http://www.unccd.int/Lists/SiteDocumentLibrary/Publications/Desertification-ENG.pdf>.
- Vallejo R, Allen E, Aronson J, et al. 2009. Restoration of Mediterranean-type woodlands and shrublands. In: van Andel J, Aronson J. *Restoration Ecology: The New Frontier* (2nd ed.). Oxford: Blackwell Publishing Ltd., 130–144.
- Villagra P E. 2000. Ecological aspects of the Argentine algarroba trees. *Multequina*, 9(2): 35–51. (in Spanish)
- Villagra P E, Giordano C, Alvarez J A, et al. 2011. To be a plant in the desert: water use strategies and water stress resistance in the central Monte desert from Argentina. *Ecología Austral*, 21: 29–42. (in Spanish)
- Weiherr E, van der Werf A, Thompson K, et al. 1999. Challenging Theophrastus: A common core list of plant traits for functional ecology. *Journal of Vegetation Science*, 10(5): 609–620.
- Westoby M, Wright I. 2006. Land-plant ecology on the basis of functional traits. *Trends in Ecology and Evolution*, 21(5): 261–268.
- Whitford W. 2002. *Ecology of Desert Systems*. San Diego: Academic Press, 97–119.
- Wright I, Westoby M. 2002. Leaves at low versus high rainfall: Coordination of structure, lifespan and physiology. *New Phytologist*, 155(3): 403–416.
- Zeberio J M. 2012. Progress of the agricultural frontier in the northeast of Patagonia and its consequences in the processes of desertification and loss of biodiversity. In: Dos Santos Afonso M R. *Environmental Science and Technology*. Buenos Aires: Argentine Association for the Progress of Science, 216–221. (in Spanish)
- Zeberio J M, Calabrese G M. 2013. Pregerminal treatments in three species of the genus *Prosopis*. In: Pérez D, Rovere A, Rodríguez Araujo M E. *Ecological Restoration in the Arid Diagonal of Argentina*. Buenos Aires: Vazquez Mazzini, 140–149. (in Spanish)
- Zeberio J M. 2018. Conservation status and rehabilitation possibilities in semi-arid ecosystems: the case of Monte in the Northeast of Río Negro. Ph.D. Dissertation. La Plata: National University of La Plata. (in Spanish)
- Zeberio J M, Torres Robles S, Calabrese G M. 2018. Land use and conservation of woody vegetation in the Northeast of Patagonia. *Ecología Austral*, 28(4): 543–552. (in Spanish)
- Zirbel C, Bassett T, Grman E, et al. 2017. Plant functional traits and environmental conditions shape community assembly and ecosystem functioning during restoration. *Journal of Applied Ecology*, 54(4): 1070–1079.